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Anaerobic degradation of organic carbon supports uncultured microbial populations in estuarine sediments

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Abstract

Background A large proportion of prokaryotic microbes in marine sediments remains uncultured, hindering our understanding of their ecological functions and metabolic features. Recent environmental metagenomic studies suggested that many of these uncultured microbes contribute to the degradation of organic matter, accompanied by acetogenesis, but the supporting experimental evidence is limited.

Results Estuarine sediments were incubated with different types of organic matters under anaerobic conditions, and the increase of uncultured bacterial populations was monitored. We found that (1) lignin stimulated the increase of uncultured bacteria within the class Dehalococcoidia. Their ability to metabolize lignin was further supported by the presence of genes associated with a nearly complete degradation pathway of phenolic monomers in the Dehalococcoidia metagenome-assembled genomes (MAGs). (2) The addition of cellulose stimulated the increase of bacteria in the phylum *Ca. Fermentibacterota* and family *Fibrobacterales*, a high copy number of genes encoding extracellular endoglucanase or/and 1,4-beta-cellobiosidase for cellulose decomposition and multiple sugar transporters were present in their MAGs. (3) Uncultured lineages in the order *Bacteroidales* and the family *Leptospiraceae* were enriched by the addition of casein and oleic acid, respectively, a high copy number of genes encoding extracellular peptidases, and the complete β -oxidation pathway were found in those MAGs of *Bacteroidales* and *Leptospiraceae*, respectively. (4) The growth of unclassified bacteria of the order *Clostridiales* was found after the addition of both casein and cellulose. Their MAGs contained multiple copies of genes for extracellular peptidases and endoglucanase. Additionally, ¹³C-labeled acetate was produced in the incubations when ¹³C-labeled dissolved inorganic carbon was provided.

Conclusions Our results provide new insights into the roles of microorganisms during organic carbon degradation in anaerobic estuarine sediments and suggest that these macro and single molecular organic carbons support the persistence and increase of uncultivated bacteria. Acetogenesis is an additional important microbial process alongside organic carbon degradation.

Keywords Estuarine sediment, Organic carbon, Anaerobic degradation, Uncultured bacteria, Acetogenesis

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Introduction

Estuaries and shallow continental shelves receive large amounts of organic carbon (OC) from terrestrial and marine sources. Consequently, nearshore marine sediments play an important role globally in organic carbon burial and diagenesis and account for ~45% of total organic carbon burial in marine sediments [1, 2]. The degradation/mineralization of OC in sediments follows the microbial respiration cascade as oxygen, nitrate, manganese, iron, sulfate, and carbon dioxide, which are subsequently used as electron acceptors with decreasing availability of Gibbs free energy [3]. In estuarine environments with high organic loading, oxygen, nitrate, manganese, iron, and sulfate are quickly exhausted within a few tens of centimeters, followed by an extensive methanogenic zone (MZ) where methanogenesis occurs predominantly [4, 5]. Microbially mediated degradation of organic carbon is a complex process during which larger macromolecules are initially broken down into monomers or oligomers that are subsequently fermented to low molecular weight intermediates such as H₂, alcohols, lactate, acetate, propionate, and butyrate. These intermediates are eventually metabolically converted to CH₄ and CO₂ [6], and methanogenesis in the MZ is the terminal step of microbial mineralization of organic carbon. In the upper MZ, a large proportion of OC [7], comprising macromolecular carbohydrates, proteinaceous compounds, aromatic compounds, and humic substances, remains undegraded and buried in the MZ [8, 9]. Most studies investigating the MZ have focused on methanogenesis and metabolism of the low molecular weight intermediates [10–13], while its upstream process about the degradation of the residual higher molecules and the responsible microbes remain undetermined.

The microorganisms in marine sediments are dominated by uncultured groups such as archaea affiliated with *Ca.* Bathyarchaeota, *Ca.* Woesearchaeota, Lokiarchaeota, and Thermoplasmatales; and bacteria affiliated with Chloroflexi, Atribacteria, *Ca.* Fermentibacterota, Planctomycetes, Clostridia, and Bacteroidetes [14–24]. However, information on the metabolic properties and functions of these above-mentioned uncultured microbes was left with little understanding. Most studies are based on the predictions following the metagenome-assembled genome (MAG) analyses, which suggests that they might play important roles in OC degradation. For example, the identified genes of different encoded enzymes associated with (1) the degradation, transport, and utilization of detrital proteins in *Ca.* Bathyarchaeota and Lokiarchaeota [25–27]; (2) the degradation of aromatic compounds in Chloroflexi, Lokiarchaeota, and *Ca.* Bathyarchaeota [28–30]; (3) the fermentation of carbohydrates in *Ca.* Fermentibacterota [23]; and (4) the anaerobic degradation of

hydrocarbon in Atribacteria [31]. To further illustrate the roles of the uncultivated microbes in the degradation of sedimentary OC and microbial interactions in the ecosystem, culture-dependent experimental evidence such as incubation or stable isotope probing approaches are needed.

Acetate is the key microbial metabolite in the carbon cycling of anoxic marine sediments. Besides macromolecular fermentation or hydrolysis, acetogenesis through CO₂ reduction is considered to be an important microbial process in anoxic sediments [32–34]. Acetogenesis is the process by which acetate is synthesized de novo through CO₂ reduction by the “Wood–Ljungdahl” (WL, reductive acetyl CoA) pathway, where various energy substrates act as electron donors, such as H₂, CO, formate, lactate, methanol, syringate, and vanillate [34]. The capability of acetogens to metabolize different types of energy substrates coupled with CO₂ reduction provides them ecological advantages [33]. For example, acetogenic microbes using methoxylated aromatic compounds (e.g., syringate and vanillate) can avoid competing for substrates in the environment with sulfate reducers and methanogens because they normally use formate, lactate, and methanol [34, 35]. The WL pathway is widely distributed in the metagenomes of marine sediments [25, 36] and also in various sedimentary archaeal and bacterial genomes [37, 38]. Incubation with ¹³C-labeled bicarbonate showed acetogenic growth of a bathyarchaeotal group with lignin as an energy source and CO₂ as a carbon source for biomass accumulation and acetate production [24]. The DNA-based stable isotope probing (DNA-SIP) and gene expression data demonstrated acetogenic activity of Lokiarchaeota, whereby fermentative H₂ production from organic substrates is coupled with the WL carbon fixation pathway [27]. The ability to degrade OC coupled to acetogenesis gives the members of *Ca.* Bathyarchaeota and Lokiarchaeota a competitive advantage to distribute globally in marine sediments. However, it remains unclear if acetogenesis is a common metabolic strategy in subsurface sediments.

Therefore, this study aims to address the following questions: (1) could the higher molecule organic compounds in the deep MZ of marine sediment be further degraded by microbes? and by which type of microbes? (2) which pathways are used by these microbes? and (3) is acetogenesis a common strategy of sedimentary microbes? We hypothesize that subsurface microbes have the capability of utilizing diverse types of organic compounds through the combination of organic carbon degradation and acetogenesis in MZ. To address these questions and test the hypothesis, we had set up a series of enrichments from estuarine sediments under anaerobic conditions that mimicked the MZ, with the

addition of diverse OC substrates and ^{13}C -labeled dissolved inorganic carbon. These OC substrates include the long-chain fatty acid oleic acid, the protein casein, the phenolic polymer lignin, and the polymeric carbohydrate cellulose, which represent commonly occurring organic matter in estuarine sediments [24]. In our previous study, we monitored the growth of uncultivated *Ca. Bathyarchaeota* following stimulation by the addition of lignin [24]. This study builds on the previous study by monitoring the response of various uncultured microbial groups as well as the intensity of acetogenesis in the above-mentioned substrate amendments.

Materials and methods

Sample collection and incubation conditions

Intertidal sediments were collected from Dayangshan island (30.592817 N, 122.083493 E) in the Hangzhou Bay of the East China Sea, and samples from 10 cm below the surface were used for incubations [24]. The samples were kept in oxygen-free gas-tight bags on ice and transported to the laboratory within 3 h, then stored at 4 °C until further processing.

The set-up of enrichment with various organic carbon substrates is described in another study [24]. Before incubation, the sediments were washed twice with NaHCO_3 -free and Na_2SO_4 -free artificial seawater medium [39], thus diluting sulfate and dissolved organic carbon (DOC) present in the original porewater. Then, the sediment was mixed again with an anaerobic artificial seawater medium (NaHCO_3 -free and Na_2SO_4 -free) and divided them equally. To test if inorganic C is transformed into acetate, 5 mM NaHCO_3 containing 5% (mol/mol) ^{13}C was added to one half of the samples, 5 mM NaHCO_3 without ^{13}C was added to the other half, and then the samples were dispensed into serum bottles as sediment slurries (100 mL). The ratio of liquid to sediment in slurries was about 10:1. The following five organic substrates were added to different experimental set-ups with a final concentration of 50 mg/L in slurries: oleic acid (Sinopharm, China), casein (Sinopharm, China), lignin (Sigma Aldrich, China), and cellulose (Sinopharm, China). Two replicates of each treatment were applied; the experimental setups without the addition of any organic substrate were used as controls. The slurries were incubated horizontally in the dark at 20 °C for 3.5 months ($t_{3.5}$) without shaking. Later, a higher concentration of organic substrate solution (the final concentration was 500 mg/L for each substrate) and 0.28 mM ^{13}C labeled NaHCO_3 was added to the slurries and incubated for another 2.5 months (t_6). At t_6 , more organic substrates (the final concentration was 500 mg/L for each substrate) were added again. Twenty milliliters of each slurry was collected after 6 (t_6) and 11 (t_{11}) months of incubation,

and the samples were centrifuged at $13,800 \times g$ for 10 min to separate the supernatant and the sediment and stored at -80 °C for DNA/RNA isolation and acetate measurements. The sample list used for DNA/RNA isolation and acetate measurements is presented in Table S1.

DNA extraction

Total DNA for amplifying the 16S rRNA genes was extracted from the ^{12}C - NaHCO_3 treatments (at t_6 and t_{11}) and the original sample using the PowerSoil DNA Isolation Kit (QIAGEN, China) (Table S1). The DNA for metagenomic sequencing was isolated from ^{12}C -dissolved inorganic carbon (DIC) treatments at t_{11} and the original sample using the SDS-based DNA extraction method [36, 40] (Table S1).

RNA extraction and reverse transcription

The total RNA was extracted from the ^{12}C -DIC treatments and the original sample using the RNeasy PowerSoil Total RNA kit (QIAGEN, China). In ^{12}C -DIC treatments, four samples collected at t_6 and t_{11} of each substrate and control were mixed and used for RNA extraction (Table S1). HiScript III 1st strand cDNA Synthesis Kit (Vazyme, China) was used to perform reverse transcription. In order to exclude contaminated sequences generated during experiments, one negative control was carried out by adding sterilized water as a template from the RNA extraction. Another negative control was carried out by adding sterilized diethylpyrocarbonate (DEPC) water as the template from the reverse transcription step. The reversed cDNA from these two negative controls were also used for the amplifying and sequencing.

Illumina sequencing and data analysis

The hypervariable V4 region of the prokaryotic 16S rRNA genes and 16S rRNA was amplified using the primer set Bac520F/Bac802R [41]. The thermal cycling program was as follows: initial denaturation at 95 °C for 4 min, 30 cycles at 95 °C for 30 s, 55 °C for 60 s, and 72 °C for 60 s, and a final extension at 72 °C for 7 min. Each reaction mixture (50 μL) contained $10 \times$ PCR buffer, dNTPs (100 μM each), 0.25 μM of each primer, 2.5 U of DNA polymerase (Ex-Taq; TaKaRa, China), and approximately 10 ng of total DNA/cDNA. PCR products were purified using the E.Z.N.A. Gel Extraction Kit (Omega Bio-Tek, Norcross, GA, USA) following the manufacturer's instructions. The sequence reads were obtained from the MiSeq Reagent Kit v2 (500 cycles, Illumina, San Diego CA, USA) and the Illumina MiSeq platform based on 2×250 bp cycles following the manufacturer's instructions. Further analysis was performed using the QIIME 2 standard pipeline [42]. For 16S rRNA diversity analysis,

the sequences in two negative controls were analyzed together with those from the experimental groups. At the species level, the groups which show high relative abundance (>1%) in negative controls were considered as contaminated sequences and manually removed from experimental groups.

Sequencing, assembly, binning, and annotation of the metagenome

For metagenomic analysis, paired-end sequencing was performed using a 2×100 bp Illumina HiSeq 2000 platform (TruSeq SBS KIT-HS V3, Illumina, at BGI-Shenzhen, China), and approximately 30G raw reads were obtained from each sample. Sickie (<https://github.com/najoshi/sickle>) was used to dereplicate and trim the raw shotgun sequencing metagenomic reads with the “pe” option and default setting. The dereplicated, trimmed, and paired-end DNA reads were assembled using MEGAHIT [43] with the following parameters: k-min 31, k-max 127, and step 4 [44]. The 200,145–561,880 assembled contigs, which were longer than 1 kb were obtained and used to binning into putative taxonomic groups based on abundance information using MaxBin version 2.2.4 with the run MaxBin.pl script [45]. The advantage of MAG-based metagenome analysis is that microbial genome information can be obtained without pure culture, and the disadvantage is that these genomes is incomplete and there are possible contamination sequences. The detailed estimates of genome contamination and completeness were assessed based on lineage-specific marker sets with CheckM [46]. The MAGs were classified based on the Genome Taxonomy Database (GTDB) (<https://gtdb.ecogenomic.org/>). All retrieved MAGs were annotated using eggnog-mapper-1.03 in the EggNOG database with e value 10^{-10} . The MAGs were assessed for the completeness of specific pathways and functions based on the canonical pathways available in the KEGG pathway database (www.kegg.jp), and the protein family databases from Pfam 31.0. SignalP [47] were used for signal peptide predictions. The abundances of MAGs across enrichment cultures and original sample metagenomes were estimated according to the method published by Pérez Castro [48]. Metagenomic reads belonging to bacterial and archaeal 16S rRNA genes were sorted using SortMeRNA v2.1b [49] and were taxonomically classified using the SILVA SSU138 database (<https://www.arb-silva.de/>). Specifically, the SILVA SSU 138 database was downloaded, and a local nucleotide database for BLASTN was constructed.

Acetate concentration and carbon isotope measurement

Before conducting the analysis, 450 μ L supernatant of each type of incubated sediment slurry was acidified

with 50 μ L of concentrated H_3PO_4 (85%) to remove ^{13}C -labeled dissolved inorganic carbon (DIC). The concentration and the stable carbon isotopic composition of acetate were analyzed by liquid chromatography coupled to isotope ratio mass spectrometry (LC-IRMS), following the method described by Heuer et al. [32]. Accordingly, the detection limit for quantitative analysis is 5 μ M and for precise $\delta^{13}C$ analysis of unlabeled samples 10 μ M [32]. Since the uptake of ^{13}C can be unambiguously detected in samples with acetate concentrations as low as 4 μ M, we also report the $\delta^{13}C$ values for samples with a concentration lower than 10 μ M (Table 1). The standard deviation of replicate $\delta^{13}C$ analysis of incubated samples ranged between ± 0.2 and 400‰ dependent on low or high ^{13}C -label, respectively, which is consistent with previous observations by Aepfler et al. [50]. The standard deviation for concentration analysis ranged from 1 to 30 μ M. Some samples were analyzed once due to sample limitation.

Results

Our previous study [24] focused only on the growth of the members of *Ca. Bathyarchaeota* stimulated by lignin. This study elucidated the growth responses of different types of sedimentary microbes to amendments of structurally diverse organic compounds. Four substrates, including oleic acid, casein, lignin, and cellulose served as model compounds for the common sedimentary constituents, i.e., long-chain fatty acids, proteins, aromatic compounds, and polymeric carbohydrates, respectively [24]. Casein and cellulose stimulated the growth of microbes, while microbial growth was not stimulated by oleic acid and lignin (Fig. S1). This was probably because the organic matter of proteins and carbohydrates facilitated the division and proliferation of microbial cells.

The shift in archaeal and bacterial communities across treatments with OC

The changes in the microbial communities were determined by high-throughput prokaryotic 16S rRNA gene amplicons, and the growth responses of some microbial groups to these amended OC were monitored. The compositions of the prokaryotic community at the phylum level in the original sediment, and after 6 and 11 months of incubation (t_6 and t_{11}), are shown in Fig. S2. Our study highlights the groups that showed growth throughout the enrichment in Fig. 1. For archaea, the relative abundance of three types of methanogens, including the members in the genus *Methanococcus*, the genus *Methanocalculus*, and the order Methanosarcinales, increased in all treatments (Fig. 1). Besides the abundance of the uncultured phylum of *Ca. Bathyarchaeota* influenced by lignin [24], the relative abundance of the uncultured phylum of *Ca.*

Table 1 Carbon isotopic composition and amount of acetate in the supernatant after incubation. Data were average with standard deviation (SD) after duplicate analysis; those without SDs were analyzed once; n.d., not determined; t_6 and t_{11} refer to the harvest time of samples after 6 months and 11 months, respectively

Conditions	^{13}C -DIC	time	$\delta^{13}\text{C}$ (‰; VPDB)	Concentration (μM)	Conditions	^{13}C -DIC	time	$\delta^{13}\text{C}$ (‰; VPDB)	Concentration (μM)
Control	NO	t_6	-8.2	6	Casein	NO	t_6	-2.2	8.6
		t_6	-10.8	5			t_6	-4.3	4.4
	YES	t_6	1740	7		YES	t_6	920	79
		t_6	1530 \pm 400	6 \pm 4			t_6	730	102
	NO	t_{11}	-28.1	20		NO	t_{11}	-9.8	1303
		t_{11}	-21.6	11			t_{11}	-9.8	1189
	YES	t_{11}	13	4		YES	t_{11}	570	1049
		t_{11}	n.d.	n.d.			t_{11}	450	979
Cellulose	NO	t_6	n.d.	n.d.	Lignin	NO	t_6	-14.5 \pm 6.5	5 \pm 2
		t_6	n.d.	n.d.			t_6	-19.9 \pm 8.2	6 \pm 2
	YES	t_6	990	65.2		YES	t_6	530 \pm 10	34 \pm 1
		t_6	1410	40.1			t_6	1990 \pm 230	5 \pm 2
	NO	t_{11}	-15.3	27.0		NO	t_{11}	-15.9	26
		t_{11}	-18.7	289			t_{11}	-17.2	44
	YES	t_{11}	560	219		YES	t_{11}	510 \pm 80	89 \pm 10
		t_{11}	540	10.5			t_{11}	1040 \pm 40	36 \pm 5
oleic acid	NO	t_6	18.0 \pm 0.4	280 \pm 30					
		t_6	19.0	131					
	YES	t_6	1500	193					
		t_6	1590	29					
	NO	t_{11}	-4.0	335					
		t_{11}	-9.2	414					
	YES	t_{11}	910	301					
		t_{11}	390	1137					

Woese archaeota increased substantially after treatment with casein and cellulose. The members of the phylum *Ca.* Woese archaeota have been suggested to have a symbiotic lifestyle as the size of their genomes is small and their central metabolic pathways are absent [51].

Regarding bacteria, the relative abundances of the groups in the genus *Fusibacter* (Phylum Firmicutes) and the genus *Syntrophotalea* (Phylum Desulfobacterota) increased in all amended and unamended treatments (Fig. 1), suggesting that they might play a less prominent role in the initial degradation steps of the added organic substrates. We also found that some uncultured bacterial groups grew in response to specific substrates. The relative abundances of bacteria in the uncultured phylum of *Ca.* Fermentibacterota and unclassified groups in the family Fibrobacterales (Fibrobacterota) increased solely after cellulose treatment. Lignin treatment specifically increased the relative abundance of unclassified lineages in the class Dehalococcoidia (Chloroflexi). Casein and cellulose both stimulated the growth of the unclassified groups in the order Clostridiales (Firmicutes). Specific

enrichments were also found in the unclassified groups of order Bacteroidales (Bacteroidota) and the family Leptospiraceae (Spirochaetota), which increased after treatment with casein and oleic acid, respectively (Fig. 1). Besides the relative abundances, the cell numbers of the abovementioned lineages of uncultured bacterial groups increased by several-fold to three orders of magnitude, calculated from the total prokaryotic 16S rRNA gene copy numbers and their relative abundance in the microbial community (Fig. S1).

Moreover, to verify the activity of these specific groups in the corresponding cultures, the transcripts of 16S rRNA genes were also detected. The results showed that the relative abundances of 16S rRNA in these incubated samples were similar to those of the 16S rRNA gene (Fig. S3). Cell transcription was detected for all the abovementioned growing uncultured bacteria, with particularly high levels of activity observed for the order Bacteroidales, family Fibrobacterota, and Leptospiraceae after treatment with casein, cellulose, and oleic acid, respectively. Metagenomic reads belonging to bacterial and

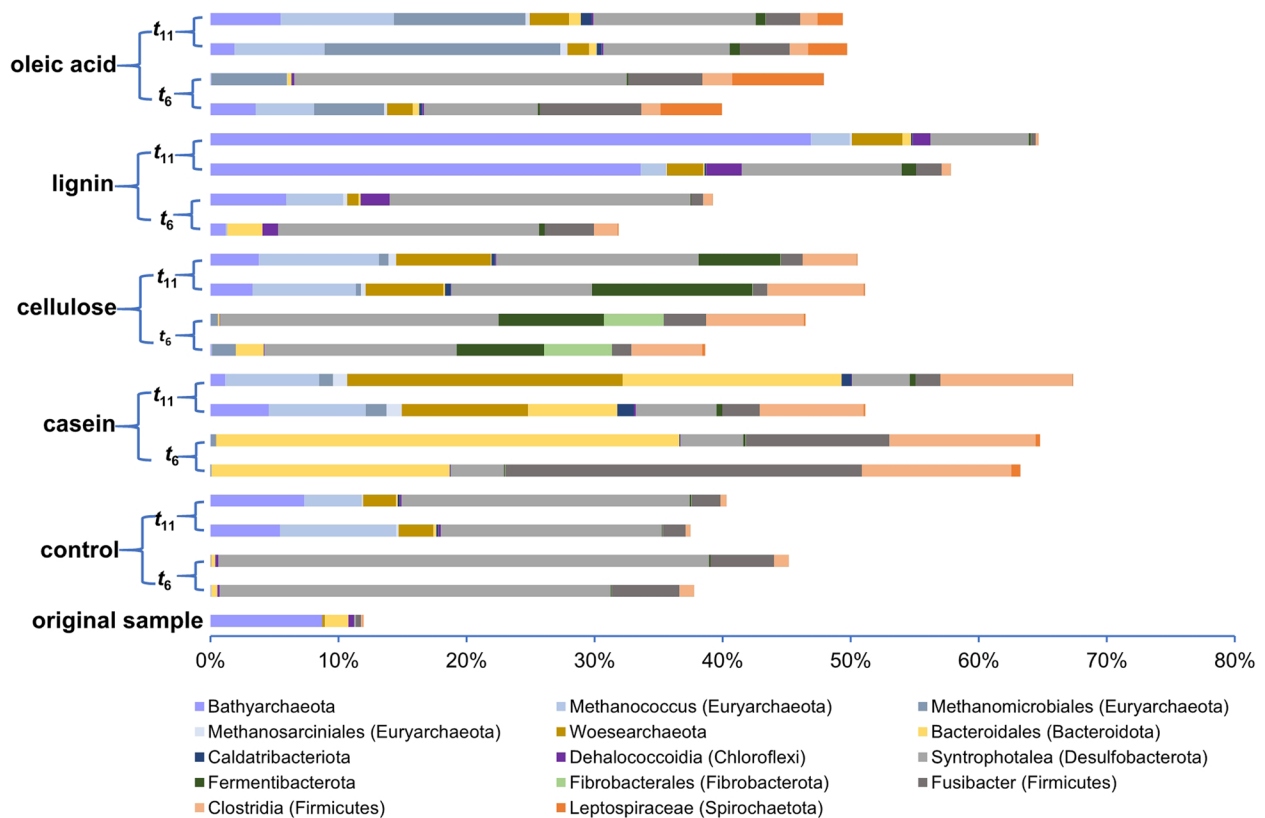


Fig. 1 The relative abundance of microbial groups showing a response to the addition of different OMs based on analysis of 16S rRNA gene amplicons. Shown is the increase in the relative abundance of *Methanococcus*, *Methanocalculus*, Methanosarciniales, *Ca.* Bathyarchaeota, *Ca.* Woeseearchaeota, *Ca.* Fermentibacterota, Fibrobacteriales, Bacteroidales, *Fusibacter*, Clostridiales, *Syntrophotalea*, Leptospiraceae, and Dehalococcoidia. t_6 and t_{11} refer to the harvest time of samples after 6 months and 11 months, respectively

archaeal 16S rRNA genes were also extracted for analysis of microbial composition (Fig. S4), and it showed a similar trend with these based on the 16S rRNA gene and 16S rRNA amplicon.

Change in abundance of genes related to OC degradation

The genes related to OC degradation were retrieved from the metagenomic data, and their abundances were quantitatively compared (Fig. S5). Genes related to the β -oxidation pathway increased after the addition of oleic acid, including those encoding long-chain fatty acid acyl-CoA synthetase, acyl-CoA dehydrogenase, enoyl-CoA hydratase, and acetyl-CoA acyltransferase. The abundances of genes encoding peptidase and protease were much higher in the amendment with casein compared to those in other amendments. Genes encoding catalase-peroxidase and benzoyl-CoA reductase (*bcr*) for lignin and aromatic compound degradation were most abundant in the amendment with lignin. Likewise, genes encoding endoglucanase and cellobiose phosphorylase, responsible for cellulose degradation, were enriched by adding cellulose. The change in abundance of genes

related to different OC amendments suggests a clear response of relevant functional microbes towards different OC inputs.

OC degradation pathways in the MAGs of the specific enriched microbes

Metagenomic data were also used for assembly and binning, the obtained metagenome-assembled genomes (MAGs) were classified based on the Genome Taxonomy Database (GTDB). MAGs of these specific enriched bacteria described above, i.e., Dehalococcoidia, *Ca.* Fermentibacterota, Fibrobacteriales, Bacteroidales, Leptospiraceae, and Clostridiales were selected (Table S2) and further analyzed. MAGs of Bacteroidales, Dehalococcoidia, and Leptospiraceae showed higher abundance in casein, lignin, and oleic acid treatment, respectively; MAGs of *Ca.* Fermentibacterota and Fibrobacteriales showed higher abundance in the cellulose treatment; MAGs of Clostridiales showed higher abundance in the casein and cellulose treatments (Table S2). The sequences of the 16S rRNA gene were extracted from most of these MAGs which showed low similarities to those from

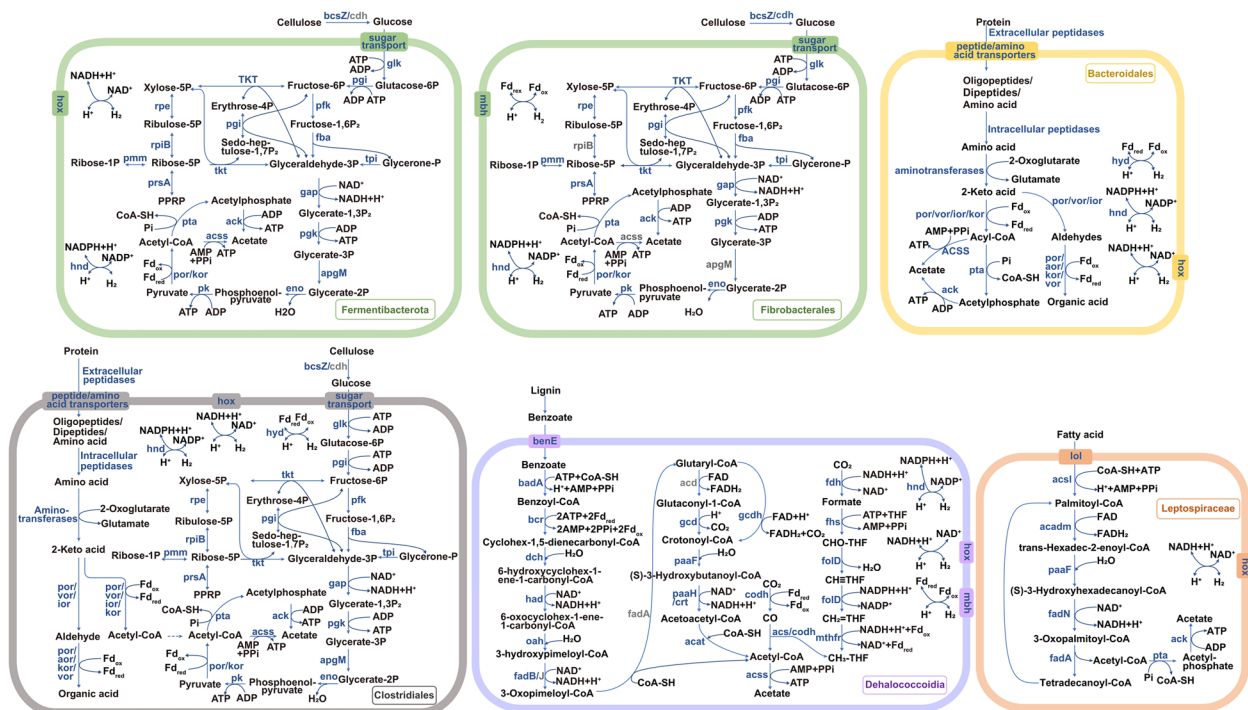


Fig. 2 Metabolic map of uncultured bacteria illustrating the degradation pathways of added substrates to acetate and H₂. Genes detected in our MAGs are shown in blue, and genes are not detected in our MAGs but are detected in the reference shown in gray. Enzymes are abbreviated with letters, and the full list as well as further metabolic pathways are provided in Supporting Information Table Tables S3, S4, S5, S6, S7 and S8

best-matching cultivated strains (between 83.3 and 92.6%) (Table S2), indicating further that these enriched bacterial groups were yet uncultivated.

Dehalococcidia

Besides the enrichment of *Ca. Bathyarchaeota* [24], we found the growth response of Dehalococcidia in lignin-amended samples (Fig. 1). Seven MAGs of Dehalococcidia were constructed (Table S2). The highest abundance of Dehalococcidia MAG contained a nearly complete benzoate degradation pathway for the anaerobic oxidation of benzoate to acetyl coenzyme A (Fig. 2 and Table S3).

Ca. Fermentibacterota and Fibrobacterales

The abundance of uncultured bacteria in the phylum *Ca. Fermentibacterota* (formerly candidate division Hyd24-12) and family *Fibrobacterales* increased specifically in the cellulose-treated sample (Fig. 1). One MAG of *Ca. Fermentibacterota* and two MAGs of *Fibrobacterales* were constructed (Table S2). The genes coding for extracellular endoglucanase or/and 1,4-beta-cellobiosidase (*cbh*), which facilitates cellulose decomposition, were present in these MAGs with high copy numbers (13–36) (Tables S4 and S5). Moreover, the genes involved in sugar transport, including the transporters of polysaccharides,

cellobiose, multiple monosaccharides, lipopolysaccharides, and maltooligosaccharides were present. The genes associated with the complete or nearly complete glycolytic and pentose phosphate pathways were also present in these MAGs (Fig. 2).

Bacteroidales

The growth of the members of Bacteroidales was specifically detected in the casein amendments (Fig. 1). Seven MAGs of Bacteroidales were obtained (Table S2), and all of them contained genes coding for oligopeptides, branched-chain amino acids, and proline transporters and high copy number of genes encoding extracellular (M23, M28, C69, S9, S41 etc.) and intracellular peptidases (M22, M23, M24, M29, M56, S46, S51 etc.) (Table S6). The genes coding for the intracellular breakdown of amino acids were also present, including ferredoxin-reducing oxidoreductases specific for aldehydes (*aor*), pyruvate/2-ketoisovalerate (*por*), indolepyruvate (*ior*), oxoisovalerate (*vor*), and 2-oxoglutarate (*kor*) (Fig. 2).

Leptospiraceae

The members of uncultured bacteria in the family Leptospiraceae were enriched specifically after the addition of oleic acid (Fig. 1). One MAG of Leptospiraceae was

constructed, and it had the genetic capability to utilize fatty acids (Fig. 2 and Table S2), i.e., it possessed genes coding for complete β -oxidation and a lipoprotein transporter (*lol*) (Fig. 2 and Table S7).

Clostridiales

The growth of uncultured groups in the order Clostridiales was detected in both casein and cellulose treatments (Fig. 1). Four MAGs of Clostridiales were constructed (Table S2). For protein and amino acid degradation, a high copy number of genes for extracellular (S9, S11, S13, S33, S41, and M23) and intracellular (M20, M24, M28, M29, M56, and S58) peptidases and genes coding for ferredoxin-reducing oxidoreductases (including *aor*, *por*, *vor*, and *kor*) were found in these MAGs of Clostridiales (Table S8). For cellulose and glucose degradation, genes coding for endoglucanase, *cbh*, and those associated with the complete or nearly complete glycolytic and pentose phosphate pathway were identified (Fig. 2 and Table S8). Additionally, the transporter genes of the branched-chain amino acid proline, lipoproteins, oligopeptides, glucose, polysaccharides, and maltose were also identified.

Acetate production by the fermentation of OC

In methanogenic sediments, it is well known that acetate can be produced by the fermentation of organic matter. Here, the production of acetate was observed in all OC treatments (Table 1). The concentrations of acetate in lignin and cellulose treatments were generally lower than that in the oleic acid and casein treatments ($p=0.0002 < 0.05$) (Table 1), which was probably because lignin and cellulose are less accessible to microbes. Because of the addition of OC at t_6 , acetate concentrations in most OC treatments increased from t_6 to t_{11} (Mann–Whitney test with one-tail, the same as below; $p < 0.05$). Specifically, the addition of casein stimulated acetate production 23-fold from t_6 to t_{11} ($p=0.014 < 0.05$). The acetate production was also detected in the controls (without OC), although their concentrations were substantially lower than in treatments with OC additions and did not differ between the time points (t_6 to t_{11}) ($p > 0.05$). The small amount of acetate present in the controls is probably due to fermentation of the remaining OC in the original sediments, and since no extra OC was added at t_6 , no increase was observed between t_6 and t_{11} .

Acetate production by acetogenesis

A substantial quantity of acetate (up to $\sim 1100 \mu\text{M}$) with positive $\delta^{13}\text{C}$ values (up to $\sim 2000\text{‰}$) was produced in all incubation lines with OC addition (Table 1). The incorporation of the ^{13}C -label from ^{13}C -DIC into acetate supports our assumption that acetogenesis, involving CO_2 reduction via the WL pathway, is a secondary

process that is interrelated with the degradation of OC. The lowered $\delta^{13}\text{C}$ -acetate values at t_{11} compared to t_6 ($P=0.01 < 0.05$) go along with increased acetate concentrations, which is likely due to an increase in the generation of non-labeled DIC through the degradation of the supplied organic substrates. The control samples without OC addition also showed high $\delta^{13}\text{C}$ -acetate values, suggesting that acetogenesis is common in estuarine sediments.

Discussion

Niches of enriched microbes involved in OC degradation in MZ

A substantial fraction of organic carbon is buried in nearshore sediments where the microbial transformation of organic carbon occurs as a key process influencing carbon flow and ultimately atmospheric oxygen and carbon dioxide concentrations [1]. In nearshore sediments with high organic matter loading, electron acceptors such as oxygen, nitrate, manganese, iron, and sulfate are quickly exhausted within the top a few millimeters to centimeters, leading to an extensive methanogenic zone (MZ) where a large proportion of organic carbon persists [4, 5]. Methanogenesis in MZ mediates 28.6% of global sub-seafloor organic matter degradation [52], while methanogens can only use low molecular weight intermediates as substrates [10–13], the transformation of residual higher molecules into low molecular weight intermediates, and the responsible microbes remain undetermined.

The dry weight of vascular plants comprises a large proportion (20–50%) of the organic carbon deposited into nearshore sediments [53, 54]. Cellulose is the most abundant of vascular plants derived organic matter followed by hemicellulose and lignin [55]. Here, the uncultured bacteria within the class Dehalococcidia were stimulated in lignin-amended samples (Fig. 1 and S1). Although the known genes related to lignin polymer depolymerization were absent in Dehalococcidia metagenome-assembled genomes (MAGs) [56], genes associated with a nearly complete degradation pathway of phenolic monomers could be found (Fig. 2 and Table S3). The catalytic processes of all the reported lignin polymer depolymerization enzymes require oxygen (aerobic environment) [56], while the mechanisms involved in the anaerobic depolymerization of lignin are unclear [57]. Therefore, there might be new and unannotated genes in the genomes of Dehalococcidia for lignin polymer depolymerization in anaerobic environments. Another possibility could be the synergistic utilization of lignin, where depolymerization of lignin polymer was conducted by other types of microbes, and the members of Dehalococcidia participated in the downstream degradation of the resulting lignin/phenolic monomers. The class

Dehalococcoidia, belonging to phylum Chloroflexi, is widely distributed in different types of environments and is usually enriched in marine subsurface sediments [30, 58]. However, only a few strains (4 species) in this group have been cultivated, leaving its metabolic properties ambiguous. In total, 31 MAGs of Dehalococcoidia were collected from the NCBI prokaryote genome database (<https://www.ncbi.nlm.nih.gov/assembly/>) and analyzed along with the MAGs obtained in this study (Table S3). Contrary to the MAGs constructed from groundwater, seawater, or wastewater, etc., most MAGs derived from marine sediments contained the benzoyl-CoA reductase (*bcr*) genes, encoding the key enzyme involved in the anaerobic degradation of aromatic compounds. Thus, the marine sediment-derived Dehalococcoidia probably play an important role in the degradation of aromatic compounds including lignin derivatives.

The addition of cellulose stimulated the increase of bacteria in the phylum *Ca. Fermentibacterota*, family Fibrobacterales, and order Clostridiales (Fig. 1 and Fig S1); a high copy number of genes encoding extracellular endoglucanase or/and 1,4-beta-cellobiosidase for cellulose decomposition and multiple sugar transporters were present in their MAGs (Fig. 2, Tables S4, S5 and S8). The members of the phylum *Ca. Fermentibacterota* are typically uncultured bacteria that are globally distributed, and they are usually found in anoxic, organic, and/or methane-rich sedimentary settings, including marine sediment [23, 59–61]. Here, all MAGs of *Ca. Fermentibacterota* (a total of 18) were downloaded from the database (Table S4); these MAGs come from different environments, including marine microbial mats, marine sediments, anaerobic digestion of organic wastes, the mouth of dolphin, soil, and wastewater. All MAGs in *Ca. Fermentibacterota* contained the endoglucanase gene and the genes associated with the complete or nearly complete glycolytic and pentose phosphate pathways. Thus, the fermentation of cellulose might be their common strategy to survive in different anoxic environments, including anoxic estuarine sediments. The members of the family Fibrobacterales were the major degraders of cellulose in the herbivore gut; historically, they were thought to only occupy mammalian intestinal tracts [62]. However, 16S rRNA genes within Fibrobacterales were also detected in landfill sites and freshwater sediments, suggesting a potential role for this family in cellulose degradation beyond the herbivore gut [62]. Here, the carrying genes coding for cellulose degradation in our MAGs of the family Fibrobacterales further confirmed this assumption and revealed the possibility of their function in estuarine sediments (Fig. 2 and Table S5).

Fatty acids are ubiquitous components in marine sediments, and oleic acid is one major fatty acid in terrestrial

plants and marine phytoplankton [63]. The members of the family Leptospiraceae were enriched by the addition of oleic acid (Fig. 1 and Fig S1), and the complete β -oxidation pathway was found in the MAG of Leptospiraceae (Fig. 2 and Table S7). Moreover, 12 available MAGs or genomes of the family Leptospiraceae were retrieved from the database; most of them were extracted from marine environments (Table S7). All these marine-derived genomes contained genes for complete β -oxidation suggesting that fatty acid degradation was their common ecological role and survival strategy. Proteins typically constitute 10% of the organic matter found in marine sediments [64]. The growth of unclassified bacteria of the order Bacteroidales and order Clostridiales was found after the addition of proteins (casein) (Fig. 1 and Fig S1), and their MAGs contained multiple copies of genes for extracellular peptidases (Fig. 2, Table S6 and S8), which implies their role in biomineralization of proteins.

Acetogenesis are important contributors of acetate in marine sediments

A substantial quantity of acetate with positive $\delta^{13}\text{C}$ values was produced in all incubation lines with/without OC addition (Table 1), suggesting that acetogenesis is key biogeochemical processes and important sources of acetate in estuarine sediments. Acetate usually accumulates in anaerobic marine sediments [13, 65–69], while it is widely believed that its major source is the fermentation of OC. Only a few studies have mentioned that acetate accumulates in marine sediments due to the reduction of CO_2 by acetogens. Heuer et al. measured the $\delta^{13}\text{C}$ values of acetate in methane-rich sediments at the northern Cascadia Margin and showed that acetogenic CO_2 -reduction/fixation can coexist with methanogenic CO_2 -reduction [32]. Moreover, Lever et al. proposed that besides fermentation, and CO_2 reduction via the WL pathway might be a source of acetate in marine sediments [33, 34]. This hypothesis was supported by the fact that the WL pathway, as one of the most important pathways for carbon fixation, is widely distributed in marine sediments [36]. It is also widespread within the archaea and bacteria [38], including cosmopolitan sedimentary microbes, such as methanogenic archaea, *Ca. Hadesarchaea*, *Ca. Theionarchaea*, *Ca. Altiarchaeales*, *Ca. Thorarchaeota*, *Ca. Bathyarchaeota*, and *Lokiarchaeota* among archaea; and Planctomycetes, Proteobacteria, Acidobacteria, Desulfobacterota, Firmicutes, and Chloroflexi among bacteria [37]. Moreover, most microbes carrying the WL pathway also have the genes of degrading a wide range of OC, and their OC degradation metabolism might be coupled with the WL carbon fixation pathway.

Under methanogenic condition, CO₂ can be used as electron acceptor and CO₂ fixation can consume the cellular reducing power produced by OC metabolism.

The occurrence of acetogenesis in marine sediments via the WL pathway is supported by our metagenomic data. For example, the addition of OC resulted in an increase in the relative abundance of genes encoding the carbon monoxide dehydrogenase/acetyl-CoA synthase complex (CODH/ACS) (subunit alpha and epsilon), the key enzyme in the WL pathway (Table S9). Besides *Ca. Bathyarchaeota* [24], the complete WL pathway and the *acss* gene/*pta-ack* pathway were also found in the MAGs of the family Desulfatiglandales (Desulfobacterota) and Dehalococcoidia (Tables S3 and S10). For the uncultured groups of Dehalococcoidia and *Ca. Bathyarchaeota*, the complete WL pathway was also frequently detected in the MAGs found in other studies [30, 36, 70]. The acetogenic growth of *Ca. Bathyarchaeota* with lignin as the electron donor for CO₂-reduction with acetate production was shown earlier [24], and the members of Dehalococcoidia probably have an acetogenic metabolism similar to that of *Ca. Bathyarchaeota*. The members of phylum Desulfobacterota are globally distributed with numerous cultured representatives, and they have historically been classified in the class Deltaproteobacteria. Cultured strains of Desulfobacterota show a preference for anoxic conditions, and many members utilize sulfate, sulfite, thiosulfate, and elemental sulfur as the electron acceptor,

with different types of OC as the electron donor [71, 72]. The WL pathway is frequently detected in the genomes of the cultured and uncultured members of Desulfobacterota [73–75]. However, in contrast to their role in sulfur reduction, their acetogenic metabolism has received little attention and has rarely been tested. The capacity to perform acetogenesis was demonstrated only in two sulfate-reducing bacteria, including *Desulfotignum phosphitoxidans* and *Dethiosulfatarculus sandiegensis*, grown in the absence of sulfate [76, 77]. In this study, the three MAGs of the family Desulfatiglandales (Desulfobacterota) carrying the genes for the complete WL pathway showed higher abundance after OC addition (Table S2). Based on our findings described above, along with the evidence provided by the carbon isotope measurements of acetate, we argue that their WL pathway might be used to fix inorganic carbon for acetate production. Taken together, our study indicates that acetogenesis through the WL pathway is an important biogeochemical process in estuarine methanogenic sediments, which has been overlooked in previous studies.

Implications of organic matter biomineralization for marine carbon cycling

Although the concentrations of hydrogen and methane were not detected in this study, their production could be inferred from the growth of methanogens in the genus *Methanococcus*, genus *Methanocalculus*, and order

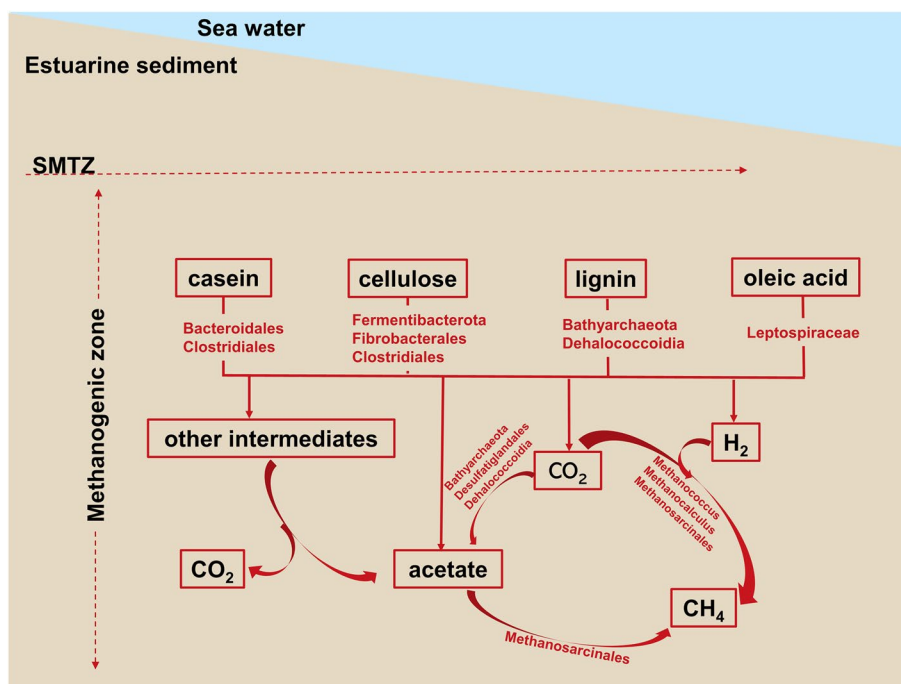


Fig. 3 Carbon flow from the added substrates to CH₄, CO₂, and acetate and the identified microbial groups involved in the cycling of these compounds in estuarine methanogenic sediments

Methanosarcinales. Methanogenesis is the terminal step of OC degradation; not surprisingly, the growth of methanogens in the different OC amending enrichments was found. The members in the genus *Methanocalculus* and *Methanococcus* are hydrogenotrophic methanogens, and the members in order Methanosarcinales are metabolically versatile methanogens, which can metabolize a wide range of substrates for methane production, e.g., H_2/CO_2 , acetate, methanol, and methylamines [78–81]. Nearly complete hydrogenotrophic pathways were found in the MAGs of these three groups (Table S11). Hydrogen also was a ubiquitous intermediate of OM fermentation in marine sediments, and most MAGs of OM degraders possessed hydrogenase genes, such as NAD-reducing hydrogenase (*hox*), NADP-reducing hydrogenase (*hnd*), and coenzyme F420 hydrogenase subunit beta (*frhB*) genes (Fig. 2, Tables S3, S4, S5, S6, S7 and S8), which implies their role in the production of hydrogen for methanogens (Fig. 3). Two MAGs classified into the order Methanosarcinales also contain the acetoclastic methanogenesis pathway (Table S11); therefore, they likely also consumed acetate for methanogenesis in these treatments (Fig. 3).

Conclusions

We conducted long-term incubations (up to 11 months) of estuarine sediments with different types of organic carbon substrates, ranging from macro to single molecules, to study the microbial processes involved to understand the associated carbon flow (Fig. 3). Based on the analysis of microbial composition (data came from 16S rRNA gene and 16S rRNA amplicons) and metabolic pathway (data came from metagenome), various uncultured bacteria showed a potential response to the degradation of the supplied carbon substrates. Furthermore, the formation of ^{13}C -labeled acetate from ^{13}C -DIC indicated that besides the fermentation of carbon substrates, microbial acetogenesis by CO_2 reduction is an important process in the estuarine sediments. For further examination of the identified organic carbon degraders, future studies should combine the metatranscriptomic analysis with the addition of isotopically labeled organic substrates for DNA/RNA-based stable isotope probing (DNA/RNA-SIP) analysis.

Abbreviations

OC	Organic carbon
IC	Inorganic carbon
MAG	Metagenome assembly genome
MZ	Methanogenic zone
DOC	Dissolved organic carbon
hox NAD	Reducing hydrogenase
hnd NADP	Reducing hydrogenase

frhB	Coenzyme F420 hydrogenase subunit beta
WL	Wood–Ljungdahl
NCBI	National Center for Biotechnology Information
KEGG	Kyoto Encyclopedia of Genes and Genomes
GTDB	Genome Taxonomy Database
DIC	Dissolved inorganic carbon
LC-IRMS	Liquid chromatography coupled to isotope ratio mass spectrometry
<i>bcr</i>	Benzoyl-CoA reductase
<i>acss</i>	Acetyl-CoA synthetase
<i>pta-ack</i>	Phosphotransacetylase-acetate kinase
<i>aor</i>	Aldehydes
<i>por</i>	Pyruvate/2-ketoisovalerate
<i>ior</i>	Indolepyruvate
<i>vor</i>	Oxoisovalerate
<i>kor</i>	2-Oxoglutarate
<i>lol</i>	Lipoprotein transporter
<i>cbh</i>	1,4-Beta-cellobiosidase
<i>CODH/ACS</i>	Carbon monoxide dehydrogenase/acetyl-CoA synthase complex
DEPC	Diethylpyrocarbonate
SD	Standard deviation
DNA-SIP	DNA-based stable isotope probing
DNA/RNA-SIP	DNA/RNA-based stable isotope probing

Supplementary Information

The online version contains supplementary material available at <https://doi.org/10.1186/s40168-023-01531-z>.

Additional file 1: Table S1. The list of samples used for DNA/RNA isolation and acetate measurements. **Table S2.** The overview of MAGs that were analyzed in this study. **Table S3.** The list of genes that are associated with benzoate degradation to acetate and H_2 production in the MAGs of Dehalococcoidia. **Table S4.** The list of genes that are associated with the cellulose degradation to acetate and H_2 production in the MAGs of *Ca. Fermentibacterota*. **Table S5.** The list of genes that are associated with the cellulose degradation to acetate and H_2 production in the MAGs of Fibrobacterales. **Table S6.** The list of genes associated with protein degradation to acetate and H_2 production in the MAGs of Bacteroidales. **Table S7.** The list of genes associated with oleic acid degradation to acetate and H_2 production in the MAGs of Leptospiraceae. **Table S8.** The list of genes associated with protein and cellulose degradation to acetate and H_2 production in the MAGs of Clostridiales. **Table S9.** The abundance of genes coding for the carbon monoxide dehydrogenase/acetyl-CoA synthase complex (*CODH/ACS*) in the metagenome data of the original sediment, control sample and treatments with different OMs at t_{11} . **Table S10.** The list of genes that are associated with the “Wood–Ljungdahl” (WL) pathway in the MAGs of Desulfatiglandales. **Table S11.** The list of genes associated with methanogenesis in the MAGs of genus *Methanococcus*, genus *Methanocalculus* and order Methanosarcinales. **Fig. S1.** The changes in the cell number of uncultured microbes in response to the addition of different OMs. The cell number was calculated from the relative abundance and prokaryotic 16S rRNA gene copy numbers. The prokaryotic 16S rRNA gene copy numbers were shown in our previous study [1]; t_6 and t_{11} indicate samples that were analyzed after 6 months and 11 months, respectively. **Fig. S2.** The comparison of prokaryotic communities at the phylum level in response to the addition of different OMs based on analysis of 16S rRNA gene amplicon; t_6 and t_{11} indicate samples that were analyzed after 6 months and 11 months, respectively. **Fig. S3.** The comparison of prokaryotic communities at the RNA level in response to the addition of different OMs based on analysis of 16S rRNA amplicon. A: The relative abundance of *Methanococcus*, *Methanocalculus*, Methanosarcinales, *Ca. Bathyarchaeota*, *Ca. Woesearchaeota*, *Ca. Fermentibacterota*, Fibrobacterales, Bacteroidales, Fusibacter, Clostridiales, Syntrophotalea, Leptospiraceae and Dehalococcoidia. B: The prokaryotic communities at the phylum level. In ^{12}C -DIC treatments, four samples collected at t_6 and t_{11} of each substrate and control were mixed and used to RNA extraction. **Fig. S4.** The comparison of prokaryotic communities in response to the addition of different OMs based on analysis of metagenomic reads. A: The relative abundance of

Methanococcus, *Methanocalculus*, Methanosarcinales, *Ca.* Bathyarchaeota, *Ca.* Woesearchaeota, *Ca.* Fermentibacterota, Fibrobacterales, Bacteroidales, Fusibacter, Clostridiales, Syntrophotalea, Leptospiraceae and Dehalococcoidia. B: The prokaryotic communities at the phylum level. In ^{12}C -DIC treatments, two samples collected at t_{11} of each substrate and control were mixed and used to DNA extraction and metagenomic sequencing.

Fig. S5. Abundance of genes involved in OC degradation in the metagenome data of original sediment and the addition of different OMs at t_{11} .

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Authors' contributions

TY and FW designed research. TY, WW and WL performed research. TY, WW, JH, YW, YC, ME, KH and FW analyzed data. TY, WW, ME, KH and FW wrote the paper. The author(s) read and approved the final manuscript.

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Availability of data and materials

Metagenomic-assembled sequences and all MAGs from the current study have been deposited in eLMSG (an eLibrary of Microbial Systematics and Genomics, <https://www.biosino.org/elmsg/index>) under accession numbers LMSG_G000011456.1-LMSG_G000011484.1. Sequences of Illumina sequencing raw data and sequences were submitted to GenBank of NCBI under accession numbers PRJNA899565.

Declarations

Ethics approval and consent to participate

Not applicable.

Consent for publication

Not applicable.

Competing interests

The authors declare no competing interests.

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References

- Berner RA. Biogeochemical cycles of carbon and sulfur and their effect on atmospheric oxygen over phanerozoic time. *Global Planet Change*. 1989;1:97–122.
- Hedges JI, Keil RG. Sedimentary organic matter preservation: an assessment and speculative synthesis. *Mar Chem*. 1995;49:123–6.
- Jørgensen BB. Mineralization of organic matter in the sea bed—the role of sulphate reduction. *Nature*. 1982;296:643–5.
- Egger M, Riedinger N, Mogollón JM, Jørgensen BB. Global diffusive fluxes of methane in marine sediments. *Nat Geosci*. 2018;11:421–5.
- Oremland RS, Taylor BF. Sulfate reduction and methanogenesis in marine sediments. *Geochim Cosmochim Acta*. 1978;42:209–14.
- Muyzer G, Stams AJ. The ecology and biotechnology of sulphate-reducing bacteria. *Nat Rev Microbiol*. 2008;6:441–54.
- Komada T, Burdige DJ, Li H, Magen C, Chanton JP, Cada AK. Organic matter cycling across the sulfate-methane transition zone of the Santa Barbara Basin California Borderland. *Geochimica Et Cosmochimica Acta*. 2016;176:259–78.
- Arndt S, Jørgensen BB, LaRowe DE, Middelburg JJ, Pancost RD, Regnier P. Quantifying the degradation of organic matter in marine sediments: a review and synthesis. *Earth Sci Rev*. 2013;123:53–86.
- Middelburg JJ. A simple rate model for organic matter decomposition in marine sediments. *Geochim Cosmochim Acta*. 1989;53:1577–81.
- Kendall MM, Liu Y, Boone DR. Butyrate- and propionate-degrading syntrophs from permanently cold marine sediments in Skan Bay, Alaska, and description of *Algorimarina butyrica* gen. nov., sp. nov. *FEMS Microbiol Lett*. 2006;262:107–14.
- Ozuolmez D, Stams AJM, Plugge CM. Propionate converting anaerobic microbial communities enriched from distinct biogeochemical zones of Aarhus Bay, Denmark under sulfidogenic and methanogenic conditions. *Microorganisms*. 2020;8:394.
- Plugge CM, Zhang W, Scholten JCM, Stams AJM. Metabolic flexibility of sulfate-reducing bacteria. *Front Microbiol*. 2011;2:81.
- Zhuang G, Xu L, Liang Q, Fan X, Xia Z, Joye SB, et al. Biogeochemistry, microbial activity, and diversity in surface and subsurface deep-sea sediments of South China Sea. *Limnol Oceanogr*. 2019;64:2252–70.
- Blazejak A, Schippers A. High abundance of JS-1- and Chloroflexi-related bacteria in deeply buried marine sediments revealed by quantitative, real-time PCR. *FEMS Microbiol Ecol*. 2010;72:198–207.
- Cui H, Su X, Chen F, Holland M, Yang S, Liang J, et al. Microbial diversity of two cold seep systems in gas hydrate-bearing sediments in the South China Sea. *Mar Environ Res*. 2019;144:230–9.
- Fry JC, Parkes RJ, Cragg BA, Weightman AJ, Webster G. Prokaryotic biodiversity and activity in the deep seafloor biosphere. *FEMS Microbiol Ecol*. 2008;66:181–96.
- Ikenaga M, Guevara R, Dean AL, Pisani C, Boyer JN. Changes in community structure of sediment bacteria along the Florida coastal everglades marsh-mangrove-seagrass salinity gradient. *Microb Ecol*. 2010;59:284–95.
- Inagaki F, Nunoura T, Nakagawa S, Teske A, Lever M, Lauer A, et al. Biogeographical distribution and diversity of microbes in methane hydrate-bearing deep marine sediments on the Pacific Ocean Margin. *Proc Natl Acad Sci USA*. 2006;103:2815–20.
- Lloyd KG, Steen AD, Ladau J, Yin J, Crosby L. Phylogenetically novel uncultured microbial cells dominate earth microbiomes. *mSystems*. 2018;3:1e00055-18.
- Orcutt BN, Sylvan JB, Knab NJ, Edwards KJ. Microbial ecology of the dark ocean above, at, and below the seafloor. *Microbiol Mol Biol Rev*. 2011;75:361–422.
- Parkes RJ, Cragg B, Roussel E, Webster G, Weightman A, Sass H. A review of prokaryotic populations and processes in sub-seafloor sediments, including biosphere:geosphere interactions. *Mar Geol*. 2014;352:409–25.
- Ruff SE, Biddle JF, Teske AP, Knittel K, Boetius A, Ramette A. Global dispersion and local diversification of the methane seep microbiome. *Proc Natl Acad Sci USA*. 2015;112:4015–20.
- Saad S, Bhatnagar S, Tegetmeyer HE, Geelhoed JS, Strous M, Ruff SE. Transient exposure to oxygen or nitrate reveals ecophysiology of fermentative and sulfate-reducing benthic microbial populations. *Environ Microbiol*. 2017;19:4866–81.
- Yu T, Wu W, Liang W, Lever MA, Hinrichs KU, Wang F. Growth of sedimentary Bathyarchaeota on lignin as an energy source. *Proc Natl Acad Sci USA*. 2018;115:6022–7.
- Lazar CS, Baker BJ, Seitz K, Hyde AS, Dick GJ, Hinrichs KU, et al. Genomic evidence for distinct carbon substrate preferences and ecological niches of Bathyarchaeota in estuarine sediments. *Environ Microbiol*. 2016;18:1200–11.
- Lloyd KG, Schreiber L, Petersen DG, Kjeldsen KU, Lever MA, Steen AD, et al. Predominant archaea in marine sediments degrade detrital proteins. *Nature*. 2013;496:215–8.

27. Orsi WD, Vuillemin A, Rodriguez P, Coskun ÖK, Gomez-Saez GV, Lavik G, Mohrholz V, Ferdelman TG. Metabolic activity analyses demonstrate that Lokiarchaeon exhibits homoacetogenesis in sulfidic marine sediments. *Nat Microbiol.* 2020;5:248–55.
28. Farag IF, Biddle JF, Zhao R, Martino AJ, House CH, León-Zayas RI. Metabolic potentials of archaeal lineages resolved from metagenomes of deep Costa Rica sediments. *ISME J.* 2020;14:1345–58.
29. Meng J, Xu J, Qin D, He Y, Xiao X, Wang F. Genetic and functional properties of uncultivated MCG archaea assessed by metagenome and gene expression analyses. *ISME J.* 2014;8:650–9.
30. Wasmund K, Schreiber L, Lloyd KG, Petersen DG, Schramm A, Stepanauskas R, Jørgensen BB, Adrian L. Genome sequencing of a single cell of the widely distributed marine subsurface Dehalococcoidia, phylum Chloroflexi. *ISME J.* 2014;8:383–97.
31. Liu Y, Qi Z, Shou L, Liu J, Yang S, Gu J, et al. Anaerobic hydrocarbon degradation in candidate phylum "Atribacteria" (J51) inferred from genomics. *ISME J.* 2019;13:2377–90.
32. Heuer VB, Pohlman JW, Torres ME, Elvert M, Hinrichs KU. The stable carbon isotope biogeochemistry of acetate and other dissolved carbon species in deep subsurface sediments at the northern Cascadia Margin. *Geochim Cosmochim Acta.* 2009;73:3323–36.
33. Lever MA. Acetogenesis in the energy-starved deep biosphere - a paradox? *Front Microbiol.* 2011;2:284.
34. Lever MA, Alperin MJ, Teske A, Heuer VB, Schmidt F, Hinrichs KU, et al. Acetogenesis in deep subsurface sediments of The Juan de Fuca Ridge Flank: a synthesis of geochemical, thermodynamic, and gene-based evidence. *Geomicrobiol J.* 2010;27:183–211.
35. Drake HL, Kuesel K, Matthies C. *Acetogenic Prokaryotes*. Edited by Dworkin M, Falkow S, Rosenberg E, Schleifer KH, Stackebrandt E. 2006. p. 354–420.
36. He Y, Li M, Perumal V, Feng X, Fang J, Xie J, et al. Genomic and enzymatic evidence for acetogenesis among multiple lineages of the archaeal phylum Bathyarchaeota widespread in marine sediments. *Nat Microbiol.* 2016;1:16035.
37. Adam PS, Borrel G, Gribaldo S. An archaeal origin of the Wood-Ljungdahl H4MPT branch and the emergence of bacterial methylotrophy. *Nat Microbiol.* 2019;4:2155–63.
38. Borrel G, Adam PS, Gribaldo S. Methanogenesis and the Wood-Ljungdahl Pathway: an ancient, versatile, and fragile association. *Genome Biol Evol.* 2016;8:1706–11.
39. Zhang Y, Henriot JP, Bursens J, Boon N. Stimulation of in vitro anaerobic oxidation of methane rate in a continuous high-pressure bioreactor. *Bioresour Technol.* 2010;101:3132–8.
40. Natarajan VP, Zhang X, Morono Y, Inagaki F, Wang F. A modified SDS-based DNA extraction method for high quality environmental DNA from seafloor environments. *Front Microbiol.* 2016;7:986.
41. Song Z, Wang F, Zhi X, Chen J, Zhou E, Liang F, et al. Bacterial and archaeal diversities in Yunnan and Tibetan hot springs. *China Environ Microbiol.* 2013;15:1160–75.
42. Bolyen E, Rideout JR, Dillon MR, Bokulich NA, Abnet CC, Al-Ghalith GA, et al. Reproducible, interactive, scalable and extensible microbiome data science using QIIME 2. *Nat Biotechnol.* 2019;37:852–7.
43. Lin X, Handley KM, Gilbert JA, Kostka JE. Metabolic potential of fatty acid oxidation and anaerobic respiration by abundant members of Thaumarchaeota and Thermoplasmata in deep anoxic peat. *ISME J.* 2015;9:2740–4.
44. Dick GJ, Andersson AF, Baker BJ, Simmons SL, Thomas BC, Yelton AP, Banfield JF. Community-wide analysis of microbial genome sequence signatures. *Genome Biol.* 2009;10:R85.
45. Wu YW, Simmons BA, Singer SW. MaxBin 2.0: an automated binning algorithm to recover genomes from multiple metagenomic datasets. *Bioinformatics.* 2016;32:605–7.
46. Parks DH, Imelfort M, Skennerton CT, Hugenholtz P, Tyson GW. CheckM: assessing the quality of microbial genomes recovered from isolates, single cells, and metagenomes. *Genome Res.* 2015;25:1043–55.
47. Petersen TN, Brunak S, von Heijne G, Nielsen H. SignalP 4.0: discriminating signal peptides from transmembrane regions. *Nat Methods.* 2011;8:785–6.
48. Pérez Castro S, Borton MA, Regan K, Hrabec de Angelis I, Wrighton KC, Teske AP, Strous M, Ruff SE. Degradation of biological macromolecules supports uncultured microbial populations in Guaymas Basin hydrothermal sediments. *ISME J.* 2021;15:3480–97.
49. Kopylova E, Noé L, Touzet H. SortMeRNA: fast and accurate filtering of ribosomal RNAs in metatranscriptomic data. *Bioinformatics.* 2012;28:3211–7.
50. Aepfler RF, Bühring SI, Elvert M. Substrate characteristic bacterial fatty acid production based on amino acid assimilation and transformation in marine sediments. *FEMS Microbiol Ecol.* 2019;95:fiz131.
51. Castelle CJ, Wrighton KC, Thomas BC, Hug LA, Brown CT, Wilkins MJ, et al. Genomic expansion of domain archaea highlights roles for organisms from new phyla in anaerobic carbon cycling. *Curr Biol.* 2015;25:690–701.
52. Bradley JA, Arndt S, Amend JP, Burwicz E, Dale AW, Egger M, et al. Widespread energy limitation to life in global subsurface sediments. *Sci Adv.* 2020;6:eaba0697.
53. Ertel JR, Hedges JI. The lignin component of humic substances: distribution among soil and sedimentary humic, fulvic, and base-insoluble fractions. *Geochim Cosmochim Acta.* 1984;48:2065–74.
54. Zeikus JG. Lignin metabolism and the carbon cycle. *Adv Microb Ecol.* 1981;5:211–43.
55. Hedges JI, Ertel JR, Leopold EB. Lignin geochemistry of a Late Quaternary sediment core from Lake Washington. *Geochim Cosmochim Acta.* 1982;46:1869–77.
56. de Gonzalo G, Colpa DI, Habib MH, Fraaije MW. Bacterial enzymes involved in lignin degradation. *J Biotechnol.* 2016;236:110–9.
57. DeAngelis KM, Sharma D, Varney R, Simmons B, Isern NG, Markillie LM, et al. Evidence supporting dissimilatory and assimilatory lignin degradation in *Enterobacter lignolyticus* SCF1. *Front Microbiol.* 2013;4:280.
58. Liu R, Wei X, Song W, Wang L, Cao J, Wu J, et al. Novel Chloroflexi genomes from the deepest ocean reveal metabolic strategies for the adaptation to deep-sea habitats. *Microbiome.* 2022;10:75.
59. Kirkegaard RH, Dueholm MS, McIlroy SJ, Nierychlo M, Karst SM, Albertsen M, et al. Genomic insights into members of the candidate phylum Hyd24-12 common in mesophilic anaerobic digesters. *ISME J.* 2016;10:2352–64.
60. Mills HJ, Martinez RJ, Story S, Sobecky PA. Characterization of microbial community structure in Gulf of Mexico gas hydrates: comparative analysis of DNA- and RNA-derived clone libraries. *Appl Environ Microbiol.* 2005;7:3235–47.
61. Schauer R, Roy H, Augustin N, Gennerich H-H, Peters M, Wenzhoefer F, et al. Bacterial sulfur cycling shapes microbial communities in surface sediments of an ultramafic hydrothermal vent field. *Environ Microbiol.* 2011;13:2633–48.
62. Ransom-Jones E, Jones DL, McCarthy AJ, McDonald JE. The Fibrobacteres: an important phylum of cellulose-degrading bacteria. *Microb Ecol.* 2012;63:267–81.
63. Wakeham SG, Lee C, Hedges JI, Hernes PJ, Peterson ML. Molecular indicators of diagenetic status in marine organic matter. *Geochim Cosmochim Acta.* 1997;61:5363–9.
64. Burdige DJ. Preservation of organic matter in marine sediments: controls, mechanisms, and an imbalance in sediment organic carbon budgets? *Chem Rev.* 2007;107:467–85.
65. Palacios PA, Snoeyenbos-West O, Loscher CR, Thamdrup B, Rotaru AE. Baltic Sea methanogens compete with acetogens for electrons from metallic iron. *ISME J.* 2019;13:3011–23.
66. Beulig F, Urich T, Nowak M, Trumbore SE, Gleixner G, Gilfillan GD, et al. Altered carbon turnover processes and microbiomes in soils under long-term extremely high CO₂ exposure. *Nat Microbiol.* 2016;1:15025.
67. Lang SQ, Osburn MR, Steen AD. Carbon in the deep biosphere. *Deep Carbon.* 2019. p. 480–523.
68. Carr SA, Jungbluth SP, Eloe-Fadrosch EA, Stepanauskas R, Woyke T, Rappe MS, et al. Carboxydrotrophy potential of uncultivated Hydrothermarchaeota from the subsurface crustal biosphere. *ISME J.* 2019;13:1457–68.
69. Zhuang G, Montgomery A, Samarkin V, Song M, Liu J, Schubotz F, et al. Generation and utilization of volatile fatty acids and alcohols in hydrothermally altered sediments in the Guaymas Basin, Gulf of California. *Geophysical Res Letters.* 2019;46.
70. Zhou Z, Pan J, Wang F, Gu JD, Li M. Bathyarchaeota: globally distributed metabolic generalists in anoxic environments. *FEMS Microbiol Rev.* 2018;42:639–55.
71. Widdel F, Bak F. Gram-negative mesophilic sulfate-reducing bacteria. In: Balows A, Trüper HG, Dworkin M, Harder W, Schleifer KH, editors. *The Prokaryotes*. New York: Springer; 1992. p. 3352–78.
72. Simon J, Kroneck PM. Microbial sulfite respiration. *Adv Microb Physiol.* 2013;62:45–117.

73. Ikeda-Ohtsubo W, Strassert JF, Kohler T, Mikaelyan A, Gregor I, McHardy AC, et al. 'Candidatus *Adiutrix intracellularis*', an endosymbiont of termite gut flagellates, is the first representative of a deep-branching clade of Deltaproteobacteria and a putative homoacetogen. *Environ Microbiol.* 2016;18:2548–64.
74. Langwig MV, De Anda V, Dombrowski N, Seitz KW, Rambo IM, Greening C, et al. Large-scale protein level comparison of Deltaproteobacteria reveals cohesive metabolic groups. *ISME J.* 2022;16:307–20.
75. Murphy CL, Biggerstaff J, Eichhorn A, Ewing E, Shahan R, Soriano D, et al. Genomic characterization of three novel Desulfobacterota classes expand the metabolic and phylogenetic diversity of the phylum. *Environ Microbiol.* 2021;23:4326–43.
76. Davidova IA, Wawrik B, Callaghan AV, Duncan K, Marks CR, Suflita JM. *Dethiosulfatarculus sandiegensis* gen. nov., sp. nov., isolated from a methanogenic paraffin-degrading enrichment culture and emended description of the family Desulfarculaceae. *Int J Syst Evol Microbiol.* 2016;66:1242–8.
77. Schink B, Thiemann V, Laue H, Friedrich MW. *Desulfotignum phosphitoxidans* sp. nov., a new marine sulfate reducer that oxidizes phosphite to phosphate. *Arch Microbiol.* 2002;177:381–91.
78. Galagan JE, Nusbaum C, Roy A, Endrizzi MG, Macdonald P, FitzHugh W, et al. The genome of *M-acetivorans* reveals extensive metabolic and physiological diversity. *Genome Res.* 2002;12:532–42.
79. Garcia JL, Patel BKC, Ollivier B. Taxonomic phylogenetic and ecological diversity of methanogenic Archaea. *Anaerobe.* 2000;6:205–26.
80. Kendall MM, Liu Y, Sieprawska-Lupa M, Stetter KO, Whitman WB, Boone DR. *Methanococcus aeolicus* sp nov., a mesophilic, methanogenic archaeon from shallow and deep marine sediments. *Int J Syst Evol Microbiol.* 2006;56:1525–9.
81. Singh N, Kendall MM, Liu YT, Boone DR. Isolation and characterization of methylotrophic methanogens from anoxic marine sediments in Skan Bay, Alaska: description of *Methanococcoides alaskense* sp nov., and emended description of *Methanosarcina baltica*. *Int J Syst Evol Microbiol.* 2005;55:2531–8.

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